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Abstract

The nurse plant phenomenon is an important form of facilitative interaction where a “nurse-plant” provides shelter from abiotic or biotic stress to a “beneficiary” plant. Plant facilitation strongly depends on nurse-plant traits such as size or age. This effect has been mostly attributed to the amelioration of abiotic conditions under larger nurse-plants. However, the effect of nurse-plant size on the overall facilitative process (quantitative and qualitative components) remains largely unexplored, particularly for increasing levels of biotic stress. Here, we investigated the effect of nurse-plant size on the quantitative (recruitment density) and qualitative (recruit shape and growth) components of plant facilitation along two contrasting herbivore-stressed environments. We measured natural oak recruits located under and outside nurse-plants of different size and age. Results show that both components of plant facilitation increased with nurse-plant size but were more pronounced at high biotic stress. Quantitatively, at high biotic stress, facilitation occurred in nurse plants of approximately half the size of those subject to low biotic stress. Interestingly, the qualitative component revealed different results depending on the ontogenetic stage of recruits, with a significant effect of nurse-plant size on large recruits (saplings) but not on small recruits (seedlings). Additionally, at higher biotic stress, more beneficiaries were found further inside the nurse-plant. Similarly, oak recruits located further inside the nurse plant showed greater plant quality. Although nurse-plant age and size were highly correlated, the spatial distribution and quality gradient of recruits suggests that nurse-plant size rather than age enhances plant facilitation in herbivore-dominated environments. We conclude that nurse-plant size plays a crucial role in plant facilitation but its net facilitative effect is strongly mediated by the level of biotic stress and the ontogeny of beneficiaries. We highlight the importance of considering both components (quantity and quality) of plant facilitation to fully understand how plant-plant interactions change at increasing levels of stress.

Keywords	Mediterranean environments; herbivory; plant ontogeny; nurse-plant age; stress gradient hypothesis; wild ungulates.
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Data will be made available on request

Highlights:

- Quantity and quality of oak recruits improved with increasing nurse-plant size
- Overall facilitation was mediated by the biotic stress and ontogeny of recruits
- Under high biotic stress, facilitation occurred at smaller nurse plant sizes
- Nurse plant size had a positive effect on quality of saplings but not on seedlings
- Sapling quality improved with increasing distance to the nurse-plant edge

1 **Title page**

2

3 **Nurse plant size and biotic stress determine quantity and quality of plant facilitation in oak**
4 **savannas**

5

6 **Running head**

7 *Nurse plant size and biotic stress determine plant facilitation*

8

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20 **ABSTRACT**

21 The nurse plant phenomenon is an important form of facilitative interaction where a “nurse-plant”
22 provides shelter from abiotic or biotic stress to a “beneficiary” plant. However, plant facilitation
23 strongly depends on nurse-plant traits such as size or age. This effect has been mostly attributed to
24 the amelioration of abiotic conditions under larger nurse-plants. However, the effect of nurse-plant
25 size on the overall facilitative process (quantitative and qualitative components) remains largely
26 unexplored, particularly for increasing levels of biotic stress. Here, we investigated the effect of
27 nurse-plant size on the quantitative (recruitment density) and qualitative (recruit shape and growth)
28 components of plant facilitation along two contrasting herbivore-stressed environments. We
29 measured natural oak recruits located under and outside nurse-plants of different size and age.

30 Results show that both components of plant facilitation increased with nurse-plant size but were
31 more pronounced at high biotic stress. Quantitatively, at high biotic stress, facilitation occurred in
32 nurse plants of approximately half the size of those subject to low biotic stress. Interestingly, the
33 qualitative component revealed different results depending on the ontogenetic stage of recruits,
34 with a significant effect of nurse-plant size on large recruits (saplings) but not on small recruits
35 (seedlings). Additionally, at higher biotic stress, more beneficiaries were found further inside the
36 nurse-plant. Similarly, oak recruits located further inside the nurse plant showed greater plant
37 quality. Although nurse-plant age and size were highly correlated, the spatial distribution and
38 quality gradient of recruits suggests that nurse-plant size rather than age enhances plant facilitation
39 in herbivore-dominated environments.

40 We conclude that nurse-plant size plays a crucial role in plant facilitation but its net facilitative
41 effect is strongly mediated by the level of biotic stress and the ontogeny of beneficiaries. We

42 highlight the importance of considering both components (quantity and quality) of plant
43 facilitation to fully understand how plant-plant interactions change at increasing levels of stress.

44

45 **Keywords:** Mediterranean environments, herbivory, plant ontogeny, nurse-plant age, stress
46 gradient hypothesis, wild ungulates.

47

48 INTRODUCTION

49 Positive interactions among plant species are common in a wide range of ecological settings
50 (Callaway, 1992; Bertness & Leonard, 1997; Tirado & Pugnaire, 2005; Soliveres et al., 2011). An
51 important form of facilitative interactions is the nurse plant phenomenon, where some plants (the
52 “nurse”) facilitate germination, establishment and development of other plant species (the
53 “beneficiary”) by providing shelter from abiotic stress, protection against biotic consumption, or
54 both (Franco & Nobel, 1988; Valiente-Banuet & Ezcurra, 1991; Brooker et al., 2008; Gómez-
55 Aparicio et al., 2008; Leiva et al., 2015).

56 The stress gradient hypothesis (SGH) predicts that the frequency of facilitation directly increases
57 with stress intensity (Bertness & Callaway, 1994; Callaway, 2007). Therefore, in stressful
58 environments, beneficiary species are often more dependent on nurse plants than in more benign
59 environments (Pugnaire et al., 1996a; Smit et al., 2008; Maestre et al., 2009; Smit et al., 2009).
60 For example, in arid and semi-arid environments, pioneering shrubs progressively change the
61 environment underneath them as they increase in size and age, facilitating the establishment of
62 less stress-tolerant species (Pugnaire et al., 1996b). Indeed, nurse-plant size influences the balance
63 of facilitative and competitive effects, with larger nurse plants supporting greater beneficiary
64 richness and biomass in both xeric and mesic sites (Tewksbury & Lloyd, 2001). These studies
65 suggest that the main cause of the increased facilitative effect of larger nurse plants was the
66 ameliorated abiotic conditions under them (Pugnaire et al., 1996b; Tewksbury & Lloyd, 2001;
67 Domínguez et al., 2015). Similarly, in herbivore-dominated environments, large nurse plants may
68 also provide better conditions against biotic stress (e.g. large herbivore stress) constituting a
69 physical barrier against browsing (Perea & Gil, 2014a). In this regard, the larger the plant size, the
70 higher the probability of the beneficiary to be separated from the nurse-plant edge where the risk

71 of being browsed is higher. Interestingly, in environments dominated by biotic stress (herbivore
72 pressure), facilitation mostly occurs through unpalatable nurse plants of sufficient size to prevent
73 consumption of the beneficiaries (Hay, 1986; Callaway et al., 2005; Smit et al., 2006; Perea & Gil,
74 2014a; Perea et al., 2016). Furthermore, in extreme conditions, biotic stress by herbivores can
75 significantly modify the sign (e.g. from competitive to facilitative) and strength of plant-plant
76 interactions (Callaway et al., 2005; Baraza et al., 2006; Smit et al., 2009; Perea & Gil, 2014a). As
77 large herbivore populations, particularly wild ungulates, have increased over the last decades,
78 browsing damage has imposed a limit on tree recruitment in open environments (Côté et al. 2004;
79 Gordon et al., 2004). Therefore, understanding the effect of nurse plant size and age at contrasting
80 levels of herbivory is of high ecological importance and deserves further research (Smit et al.,
81 2007; Gómez-Aparicio et al., 2008; Perea & Gil, 2014b).

82 Plant facilitation can be evaluated by quantitative and qualitative components, similar to other
83 biotic interactions (*sensu* Schupp et al., 2010). The quantitative component of plant facilitation
84 refers to the number of recruits successfully established under the nurse plant, whereas the
85 qualitative component refers to the performance and growth once plants have established. Most
86 studies have focused on either qualitative or quantitative facilitation but have very rarely
87 incorporated both (see Maestre et al., 2005). As stress conditions may change across different
88 recruit ontogenetic stages (Armas & Pugnaire 2009), results on the qualitative and quantitative
89 components of plant facilitation may vary over time. Furthermore, nurse-plant size may also affect
90 quantitative and qualitative components of plant facilitation differently as larger plants create
91 greater shelter that increases recruit density (Pugnaire et al., 1996b; Tewksbury & Lloyd, 2001),
92 but favors stronger competition with the underneath beneficiaries (Schwinning & Weiner, 1998;
93 Valladares & Pearcy, 2002). Similarly, beneficiary plants located further inside the nurse plant

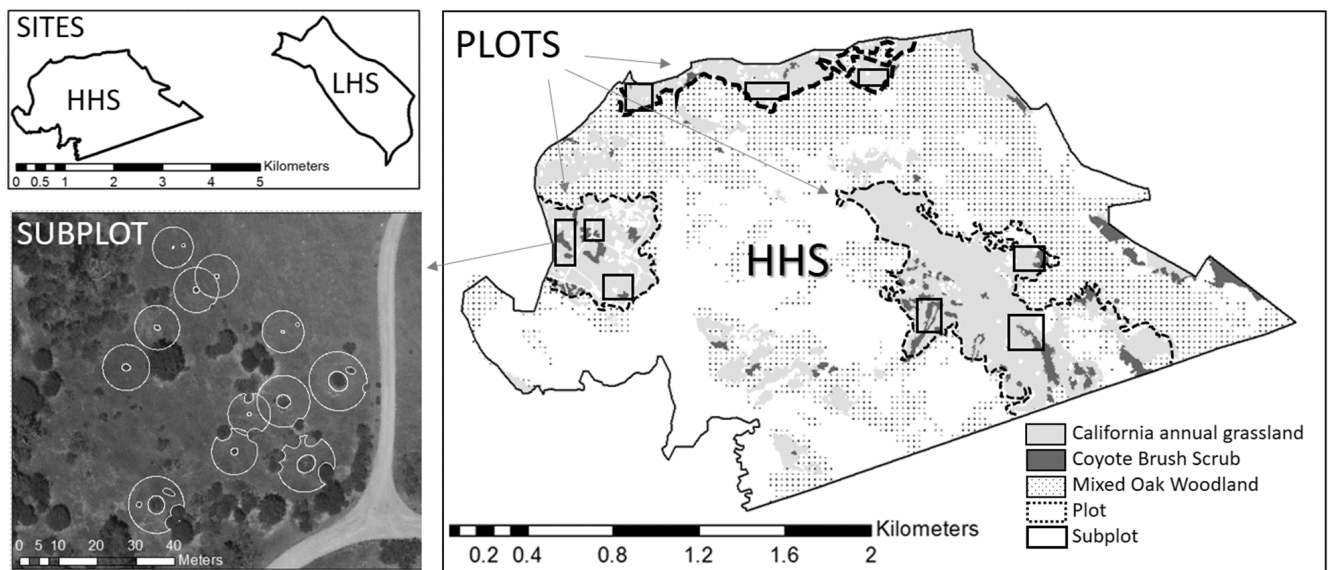
94 (i.e. at farther distances from the nurse-plant edge) will benefit from greater protection from both,
95 abiotic and biotic stress, and are therefore expected to receive stronger facilitative effects under
96 larger nurse plants. However, to the best of our knowledge, no study has fully evaluated the effect
97 of nurse-plant size on the magnitude (quantity and quality components) of the facilitative process
98 at contrasting levels of biotic stress.

99 In this study we aimed to investigate the effect of nurse-plant size on the quantitative (recruitment
100 density) and qualitative (recruit shape and growth) components of plant facilitation along two
101 contrasting herbivore-dominated environments, with high and low densities of wild ungulates. We
102 specifically focused on the facilitative relationship between a well-known nurse shrub (*Baccharis*
103 *pilularis* DC) and the beneficiary coast live oak (*Quercus agrifolia* Neé), which strongly depends
104 upon *Baccharis* cover to regenerate in California oak savannas (Callaway & Davis, 1998; Zavaleta
105 & Kettle, 2006; Perea et al., 2017). We hypothesize that both quantitative and qualitative
106 components of plant facilitation will increase with nurse-plant size but will be strongly mediated
107 by biotic stress (herbivore pressure). Specifically, we tested the following predictions: (1) overall
108 plant facilitation (both quantity and quality) will increase at higher levels of biotic stress (herbivore
109 stress) following the SGH framework, (2) nurse plants will facilitate greater density of recruits
110 with increasing plant size; (3) the qualitative component of plant facilitation will also increase with
111 both nurse-plant size and distance to the nurse-plant edge; and (4) facilitative effects due to
112 increasing nurse-plant size will be proportionally greater at higher biotic stress.

113 **MATERIALS AND METHODS**

114 *Study sites*

115 We selected two nearby sites (distance < 3 km; Fig. 1) located in the interior foothills of Northern
116 California (USA) based on their different level of biotic (herbivore) stress: a) Jasper Ridge
117 Biological Preserve, with high deer abundance and, hence, high herbivore stress (hereafter, HHS)
118 and b) the Stanford University Dish Area, with comparatively lower herbivore stress (hereafter,
119 LHS). Previous estimations showed 5.4-fold greater deer abundance in HHS than in LHS (Perea
120 et al., 2017).



121
122 **Fig. 1.** Sampling design with Plots and Subplots nested within the fixed factor Site. Each site (HHS and
123 LHS) contained three large grassland areas (Plots) with presence of scattered Coyote brush individuals.
124 Within each plot, three subplots were sampled. Twelve isolated *Baccharis* individuals (bottom left image)
125 were selected from each subplot based on satellite imagery, four of each size category: large, medium and
126 small. A total of 216 individuals of *Baccharis* (2 sites x 3 plots x 3 subplots x 12 individuals) were selected.

127
128 The HHS site (37.40°N, 122.23°W) is a 480-ha research preserve located in the foothills of the
129 Santa Cruz Mountains. The LHS site (37.41°N, 122.18°W) is located east from the HHS site and
130 covers 560 ha (Fig. 1). Both sites have a coastal Mediterranean climate, with warm, dry summers

131 and cool, wet winters. Mean annual precipitation is 605 mm and average daily temperatures range
132 from a minimum of 2 to 4 °C in winter to a maximum of 25 to 27 °C in summer (Zavaleta &
133 Kettley, 2006). Oak-dominated systems (woodlands and savannas) are the most common
134 vegetation types in both sites. Oak systems are dominated by an evergreen oak tree, coast live oak
135 (*Quercus agrifolia* Née), and two deciduous oak species: valley oak (*Q. lobata* Née) in deeper
136 soils, and blue oak (*Q. douglasii* Hook. & Arn.) in shallower, drier soils. Climate, vegetation (oak-
137 dominated systems) and historical land use are similar for both study sites (Palmer, 2003).
138 Livestock (mostly cattle) was removed from HHS site in 1973, and from LHS (study sites) in
139 1982-1988. Mule deer (*Odocoileus hemionus*) is the main browser and the only ungulate present
140 in both study sites. Western scrub jays (*Aphelocoma californica*) are considered the main avian
141 acorn dispersers in California oak woodlands (Carmen, 2004; Pesendorfer & Koenig, 2016). Tree
142 squirrels (*Sciurus* spp.) are the main terrestrial acorn dispersers in the study area (Perea et al.,
143 2018).

144 *Study species*

145 Coast live oak (*Quercus agrifolia* Née) is an evergreen long-lived tree that grows in lower-
146 elevation oak woodlands of California (Griffin, 1977). Coyote brush (*Baccharis pilularis* DC) is
147 an evergreen, much-branched shrub that grows in shrub communities seral to foothill woodlands
148 of California (Steinberg, 2002). Coyote brush is a colonizer, short-lived and light-demanding shrub
149 species (Hobbs & Mooney, 1986; Zavaleta & Kettley, 2006) that facilitates the recruitment of *Q.*
150 *agrifolia*, particularly in open woodlands and oak savannas (McBride, 1974; Callaway & Davis,
151 1998; Zavaleta & Kettley, 2006). Coyote brush is considered a species of very low palatability and
152 nutritious value for grazers and browsers (McBride & Heady, 1968; Smither-Kopperl, 2016),
153 whereas *Q. agrifolia* has been rated fair to poor for mule deer (Sampson & Jespersen, 1963).

154 *Study design*

155 We used a sampling design with Plots and Subplots nested within the fixed factor Site (Fig. 1). For
156 each site (HHS and LHS), we selected three large grassland areas with presence of scattered
157 Coyote brush individuals (plots) and, within each of them, three subplots were sampled (Fig. 1).
158 From satellite imagery, twelve isolated *Baccharis* individuals were selected within each subplot,
159 four of each size category: large (mean shrub diameter >3.5 m), medium (mean diameter 1.5-3.5
160 m) and small (mean diameter <1.5 m). Therefore, a total of 216 individuals of *Baccharis* (2 sites
161 x 3 plots x 3 subplots x 12 individuals) were selected and their GPS coordinates registered.
162 *Baccharis* were selected so that at least 80% of their surrounding area was covered by grass cover
163 (open microsite) to ensure “edge effects” and statistical independence. Subplot size was
164 approximately 500-900 m², with homogeneous ecological conditions (soil type, topography and
165 vegetation).

166 In the field, the total nurse-plant canopy cover for each individual was calculated based on its
167 perpendicular crown diameters. We also recorded the percentage of the cover that was efficiently
168 protecting the shrub as the shape of the shrub was highly variable. Beneath the canopy of each
169 sampled nurse plant and in a 5 meters concentric area surrounding it, we recorded all young oaks
170 (< 5 cm DBH, diameter at 130 cm height) and measured: 1) height and basal diameter to the nearest
171 millimeter; 2) herbivory damage produced by mule deer distinguishing between 6 different
172 browsing intensities in a 0-5 scale, following Perea et al., (2015): 0=No browsing evidence;
173 1=Light browsing, with <10% twigs browsed; 2=Low browsing, with 10-30% of twigs browsed;
174 3=Intense browsing (30-60% of the twigs); 4=Heavy browsing (>60% of the twigs), with clear
175 modification of plant shape; 5=Maximum browsing (>90% of the twigs browsed); 3) microsite
176 location (under the nurse plant vs. open), and 4) distance to the nurse-plant edge in cm. Young

177 oaks were classified in two categories: seedlings (basal diameter <10 mm) and saplings (basal
178 diameter ≥ 10 mm) following previous studies in oak savannas (López-Sánchez et al., 2014; Perea
179 et al., 2016). In addition, groups of deer pellets were also counted in the 5 meters concentric area
180 outside the nurse plant to control for differences in herbivory pressure between plots. Finally, we
181 measured the distance from each nurse plant to the nearest adult *Quercus agrifolia* using a
182 handheld laser distance meter (Leica DISTO™).

183 To calculate the density of young oaks under the nurse plant, we divided the number of young oaks
184 by the area efficiently protected by the nurse plant (area of perpendicular crown diameters
185 corrected by the shape of the nurse plant). Density of young oaks in the 5 m concentric area outside
186 the nurse plant was also calculated. To calculate the area outside the target nurse plant, we
187 subtracted any other type of cover (e.g. rock cover, small shrubs) to only account for grasslands.
188 The total area of each nurse plant and its 5 meters surrounding was calculated with GIS software
189 (ArcGis 10.3).

190 Finally, based on high-resolution historical aerial imagery available in the HHS site, we estimated
191 the minimum age of each *Baccharis* to assess the relationship between nurse-plant size and age.
192 We defined the minimum age of a *Baccharis* as the number of years elapsed from the first time
193 each target shrub was detected in the historical aerial imagery. All nurse shrub individuals were
194 easily detected through aerial imagery in their first years (3-5 years of age; Zavaleta & Kettley,
195 2006) since all these plants established in open microsites and were, therefore, highly visible. Thus,
196 age was estimated as the first date that an individual appears in an aerial photograph plus 4 years,
197 following the procedures of Zavaleta & Kettley (2006).

198 *Statistical analysis*

199 All statistical analyses were performed using the R statistical software R (Version 3.4.3; R Core
200 Team, 2018). Each *Baccharis* individual (under shrub) and the 5-m surrounding area (open
201 microsite) of each target shrub was the study unit. We first analyzed whether there were differences
202 in overall young oak densities (quantity component) between microsities (open vs. under nurse
203 plant) and sites (HHS vs. LHS). The response variable was oak recruit density calculated as the
204 number of young oaks per hectare, obtaining two values for each target *Baccharis* (density under
205 the nurse plant and density in the open, outside the nurse plant). We used zero-inflated distributions
206 as tree recruitment is often characterized by an excess of zero count (Gómez-Aparicio et al., 2009;
207 López-Sánchez et al., 2016). Therefore, the number of recruits per hectare was assumed to follow
208 a zero-inflated Poisson (ZIP) distribution (Lambert, 1992). Hence, we modelled oak recruitment
209 density as the result of two elements: 1) the probability of recruitment occurrence with the distance
210 to the nearest adult coast live oak tree using a binomial model and 2) the differences in oak recruit
211 density among microsities and herbivore stress (HHS vs. LHS), using a Poisson model. We used
212 the function “zeroinfl” from the R package “pscl”. Afterwards, we used the Vuong test (Vuong,
213 1989) to compare the fit of zero-inflated model with an ordinary Poisson regression.

214 We analyzed occurrence of nurse-plant facilitation using a Generalized Linear Mixed Model
215 (GLMM), with binomial error structure and logit link using the function glmer of the R package
216 “lme4”. We assumed that the nurse plant was facilitating oak regeneration when oak recruit density
217 under the target shrub was higher than in the surrounding 5 m concentric area. The response
218 variable was binary (value of 1 when density of young plants under the nurse plant exceeded the
219 oak density in the surrounding area, and 0 in the opposite situation). Individuals with no recruits
220 at all (under and outside the shrub cover) were discarded from the analysis. Predictors in the model
221 were nurse-plant size (mean crown diameter), herbivore stress (HHS vs. LHS) and their

222 interaction. Subplots nested within plots were introduced as random effect. In addition, to analyze
223 the effect of nurse-plant age on oak facilitation (only in HHS site), we fitted a polynomial term
224 (square) to model the non-linear relationship between nurse-plant size and age. Plot and subplot
225 were also included as random factors.

226 To examine the effect of nurse-plant size on oak recruit quality, we first calculated the height to
227 diameter ratio (H/D) for each young oak. We used H/D ratio as a proxy of plant growth quality
228 because H/D is considered an important measure of stability and proper plant development in
229 woody plants (Kamimura & Shiraishi, 2007; Schelhaas, 2008; Vospernik et al., 2010). Thus, in
230 herbivore-dominant environments, young plants with lower H/D ratio indicate heavier browsing
231 and inadequate development (e.g., more compact and brush-shaped; Martínez & López-Portillo,
232 2003) and, therefore, their transition to adult stages is retarded or suppressed (Bartolome et al.,
233 2002; López-Sánchez et al., 2016).

234 To analyze whether there were differences in oak growth quality among microsites (under nurse
235 plant vs. open) and herbivore stress (HHS vs. LHS), we performed a Linear Mixed Model (LMM)
236 with H/D as the response variable and plot and subplot as random structure. Predictors were
237 microsite, herbivore stress level and their interaction. Then, we calculated the relationship between
238 H/D ratio and browsing damage to assess whether changes in H/D ratio could be attributed to
239 herbivory damage or not. We performed this second model because higher H/D inside the nurse
240 plant might also be attributed to an elongation response of the plant due to reduced insolation.
241 Hence, we only attributed significant differences in plant growth quality to deer herbivory when
242 H/D showed a significant relationship with browsing damage.

243 In addition, to analyze the effect of nurse-plant size on the recruit quality we performed a LMM
244 with H/D as response variable. Predictors were nurse-plant size, microsite (under nurse plant vs.

245 open) and herbivore stress (HHS vs. LHS) and their two-way interactions (fixed factors). Subplots
246 nested within plots were introduced as random effects. Similarly, we built another LMM with H/D
247 as response variable and the following predictors: distance to the nurse-plant edge, herbivore stress
248 and microsite and their two-way interactions.

249 All possible models derived from the maximal models described above were ranked based on the
250 weighting provided by the Akaike information criterion (AIC) using the function “dredge” from
251 the R package “MuMIn”. When there were multiple candidate models with $\Delta AIC < 2$, we selected
252 the best-fitting model with the least number of parameters following the principle of parsimony
253 (Burnham & Anderson, 2002). Diagnostic plots were performed to detect heteroscedasticity in the
254 models and qqplots to detect skewness in the model residuals. Finally, predictions were obtained
255 using the “predict” function from the package “stats”.

256 **RESULTS**

257 Deer pellet densities were higher in HHS plots (243, 293 and 341 groups/ha) than in the LHS plots
 258 (26, 42, 54 groups/ha). This polarized results justified the use of Site as a proxy for herbivore stress
 259 (HHS and LHS).

260 *Overall quantitative component of the facilitative process*

261 The zero-inflated model showed significant differences in young oak density between microsites
 262 (open vs. under the nurse plant; Table 1) and herbivore stress (HHS vs. LHS; Table 1), with an
 263 overall decrease in the recruit density at increasing distances from the nearest adult tree (Table 1).
 264 For both sites (HHS and LHS), young oak density was significantly higher under the nurse plant
 265 than in the open microsite. However, the difference in oak density between microsites was more
 266 accentuated in the HHS than in the LHS (25 vs. 19 times greater, respectively).

267 **Table 1.** Summary of the Zero-inflated Model fitted to analyze the differences in density of young
 268 oak seedlings (n° of recruits/ha) between microsites and sites with different herbivore stress (HHS
 269 vs. LHS). Firstly, the probability of recruitment occurrence with the distance to the nearest adult
 270 oak tree was modelled using a binomial modeled (below) and secondly the differences in oak
 271 recruit density among microsites and herbivore stress (HHS vs. LHS) was modeled using a Poisson
 272 model (above).

<u>Count model coefficients (Poisson with log link)</u>				
Predictors	<i>Coeff.</i>	<i>SE</i>	<i>z-value</i>	<i>P</i>
Intercept	8.432	0.004	2355.77	<0.001
MICROSITE				
Open (against under nurse plant)	-2.930	0.010	-296.29	<0.001
HERBIVORE STRESS				
High herbivore stress (HHS) against LHS	-0.106	0.005	-23.17	<0.001
MICROSITE × HERBIVORE STRESS				
Open × HHS	-0.302	0.019	-16.10	<0.001
<u>Zero inflation model coefficients (binomial family with logit link)</u>				
Predictors	<i>Coeff.</i>	<i>SE</i>	<i>z-value</i>	<i>P</i>
Intercept	0.482	0.187	2.581	0.010
Distance to adult tree	0.021	0.007	2.918	0.004

275 *Overall qualitative component of the facilitative process*

276 We found that seedlings and saplings were browsed by deer differently, being saplings more
277 heavily browsed than seedlings when located in the open microsite (mean browsing damage \pm SD:
278 3.39 ± 0.92 and 0.65 ± 1.09 , respectively). Therefore, we decided to analyze seedlings (n=209)
279 and saplings (n=97) separately.

280 Seedlings at the higher herbivore stress site (HHS) showed no significant differences in height to
281 diameter ratio (H/D) between microsites ($t=-1.56$, $p=0.120$; Fig. S1a) whereas at lower herbivore
282 stress (LHS site), seedlings presented significantly higher H/D ratio in the open than under the
283 nurse plant ($t=2.04$, $p=0.0431$; Fig. S1a). However, this difference in H/D could not be attributed
284 to herbivory stress because the relationship between browsing damage and H/D in seedlings was
285 not significant in the LHS site ($t=-1.09$, $p=0.278$; Fig. S1b).

286 Conversely, for saplings, at both levels of herbivore stress (HHS and LHS) we found significantly
287 higher H/D ratios under the nurse plant than in the open (Fig. S1c). In addition, differences in H/D
288 between microsites were stronger in the HHS site ($t=-6.40$, $p<0.001$; 2.2-fold greater H/D ratio
289 under the nurse plant) than in the LHS site ($t=-2.46$, $p=0.016$; 1.36-fold greater H/D ratio under
290 the nurse plant; Fig S1c). Finally, the significant relationship between browsing damage and H/D
291 ratio at both sites ($t=-9.92$, $p<0.001$) confirmed that the decrease in H/D ratio for saplings was
292 correlated with an increase of browsing damage (Fig. S1d).

293 *Quantitative differences in facilitation depending on nurse-plant size and distance to nurse-plant*
294 *edge at contrasting levels of herbivory stress*

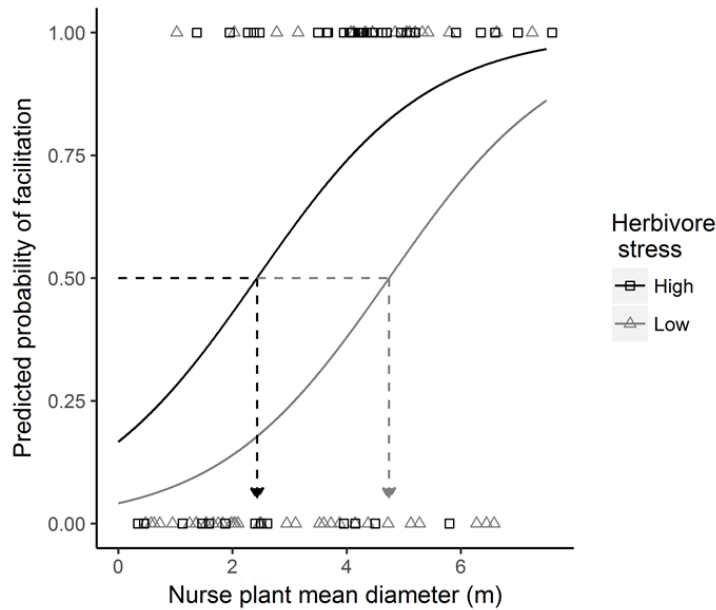
295 Occurrence of plant facilitation (i.e., the probability of having higher oak density under the nurse
296 plant than in the open) strongly increased with increasing nurse-plant size ($t=3.87$; $p<0.001$; Table
297 2; Fig. 2).

Table 2. Summary of the selected binomial GLMM to analyze the probability of plant facilitation in relation to nurse-plant size (mean crown diameter) and its variation between sites with different herbivore stress levels (HHS vs. LHS). Response variable was binary (value 1 when density of young plants under the nurse plant exceeded the oak density in the surrounding area, and 0 in the opposite situation).

Explanatory variable	Estimates	Standard error	<i>t</i> -value	<i>P</i>
Intercept	-3.15	0.80	-3.92	<0.001
NURSE-PLANT SIZE	0.66	0.17	3.87	<0.001
HERBIVORE STRESS				
High herbivore stress (HHS) against LHS	1.53	0.55	2.81	0.005

298 Bold type indicates a significant effect ($P < 0.05$).

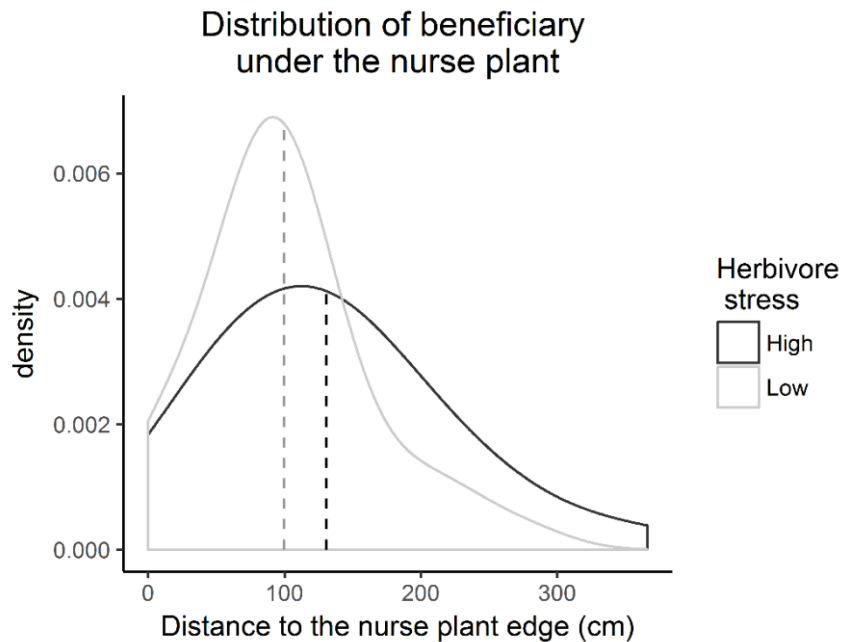
299 Furthermore, the effect of nurse-plant size on the quantitative component of plant facilitation also
 300 varied with herbivore stress (Fig. 2). Thus, the threshold size at which the nurse plant started to be
 301 an efficient facilitator (probability of facilitation > 0.5 ; Fig. 2) was approximately half at high
 302 herbivore stress (HHS site) as compared to LHS site (ca. 2.4 vs. 4.7 meter nurse-plant mean
 303 diameter; Fig. 2).



304

305 **Fig. 2.** Predicted probability of plant facilitation as a function of nurse-plant size (mean crown diameter) at
306 high (HHS; black lines) and low (LHS; grey lines) levels of herbivory. Arrows represent the threshold
307 nurse-plant size at which facilitation probability is 50%.

308 Similarly, distribution of beneficiaries underneath the nurse plant depended on the herbivore stress
309 (Fig. 3). Thus, at low herbivore stress (LHS site), recruits were found significantly closer to the
310 nurse-plant edge as compared to those located at the HHS site ($t=-2.19$ $p=0.030$; Fig. 3).



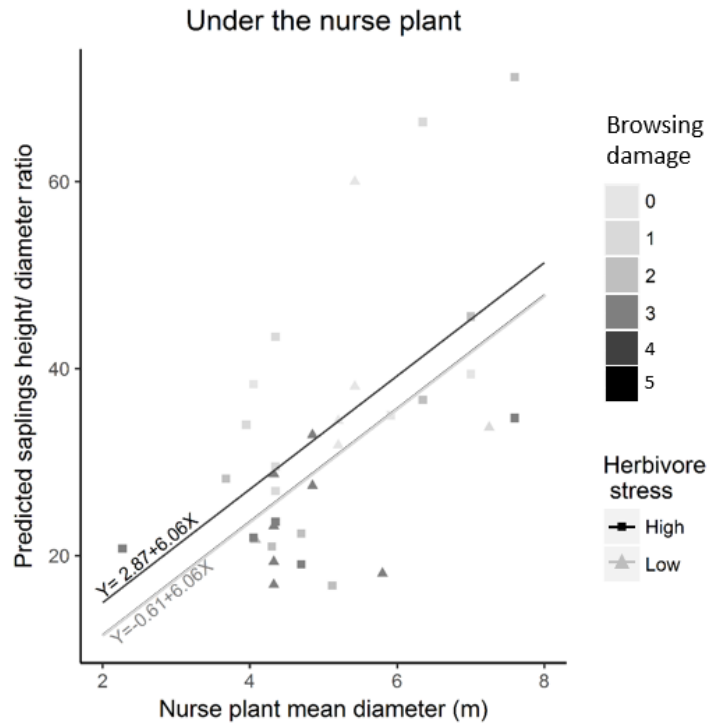
311

312 **Fig. 3.** Kernel distribution of recruits (seedlings and saplings) under the nurse plant for high (HHS) and low
313 (LHS) herbivore stress. Distribution of beneficiaries is shown in relation to distance to the nurse-plant edge.
314 Dashed lines indicate the beneficiary mean distance to the edge for HHS and LHS.

315 *Qualitative differences in facilitation depending on nurse-plant size and distance to nurse plant*
316 *edge at contrasting levels of herbivory stress*

317 Nurse-plant size was positively associated with plant quality (H/D ratio) of saplings (Table 3; Fig.
318 4), regardless of the herbivore stress level.

319 Therefore, under the nurse plant, for each meter increase in nurse-plant diameter, we observed a
320 6.06 units increase in H/D ratio ($t= 4.81$; $p<0.001$; Table 3). In contrast, nurse-plant size had no
321 effect on plant quality in the open microsite ($t=0.38$, $p=0.704$).



322

323 **Fig. 4.** Predicted H/D ratio for saplings located under the nurse-plant microsite in relation to nurse-plant
 324 size at high (HHS; black lines) and low (LHS; grey lines) herbivore stress. Square and triangles represent
 325 measured H/D ratio of saplings in HHS and LHS respectively. Color scale represents browsing damage
 326 recorded for each individual sapling.

Table 3. Summary of the selected GLMM to analyze sapling growth quality (H/D ratio) as a function of nurse-plant size (mean crown diameter), microsite (open against nurse plant) and sites with different herbivore stress levels (HHS against LHH).

327

Explanatory variable	Estimate	Standard error	t-value	p-value
Intercept	-0.61	6.80	-0.09	0.928
NURSE-PLANT SIZE	6.06	1.26	4.81	<0.001
MICROSITE				
Open against nurse plant	23.36	7.13	3.28	0.002
HERBIVORE STRESS				
High herbivore stress (HHS) against LHH	3.48	3.20	1.09	0.337
NURSE-PLANT SIZE × MICROSITE				
Nurse-plant size × Open	-6.29	1.40	-4.48	<0.001
HERBIVORE STRESS × MICROSITE				
HHS × Open	-9.89	3.84	-2.57	0.012

328

Bold type indicates a significant effect (P<0.05).

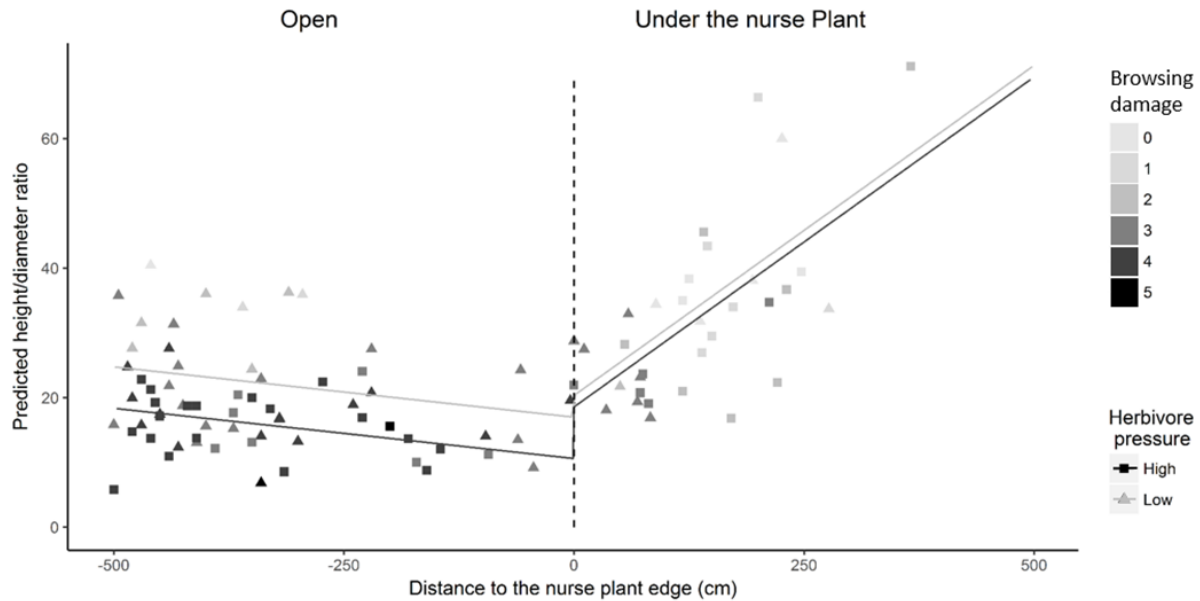
329 In addition, H/D ratio of saplings growing underneath the nurse plant was positively correlated
 330 with distance to the nurse-plant edge (H/D (under nurse plant) = $0.10 \times$ distance to nurse-plant
 331 edge (std. error = ± 0.008); $t=5.63$, $p<0.001$; Table 4; Fig. 5).

Table 4: Summary of the selected GLMM to analyze differences in H/D ratio of saplings in relation to the distance to the nurse-plant edge, microsite (open vs. nurse plant) and sites with different level of herbivore stress (HHS vs. LHS).

Explanatory variable	Estimates	Standard error	<i>t</i> -value	<i>p</i> -value
Intercept	47.52	4.06	11.71	<0.001
DIST. TO EDGE	0.10	0.02	5.63	<0.001
MICROSITE				
Open against nurse plant	-26.40	4.13	-6.39	<0.001
HERBIVORE STRESS				
High herbivore stress (HHS)	-1.79	3.57	-0.50	0.643
DIST. TO EDGE \times MICROSITE				
DIST. TO EDGE \times Open	-0.09	0.02	-4.35	<0.001
SITE \times MICROSITE				
Open \times HHS	-4.60	3.74	-1.23	0.222

332 Bold type indicates a significant effect ($P<0.05$).

333 Thus, saplings located further inside the nurse plant showed greater H/D ratios. As expected, H/D
 334 ratios of saplings located in the open microsite showed no correlation with distance to nurse-plant
 335 edge (H/D (open) = $0.016 \times$ dist. nurse-plant edge (std. error = ± 0.008); $t=1.93$, $p=0.057$; Fig. 5).
 336 Additionally, we found significant differences in sapling quality between HHS and LHS in the
 337 open microsite, while these differences were rendered under the nurse plant (Table 4; Fig 5).



338

339 **Fig. 5.** Predicted H/D ratio for saplings in relation to the distance to the nurse-plant edge for both microsites
 340 (open and under nurse plant) and for high (HHS) and low (LHS) herbivore stress. Square and triangles
 341 represent measured H/D ratio of saplings in HHS and LHS, respectively. Color scale represents browsing
 342 damage recorded for each individual sapling.

343 Seedlings performed differently in terms of plant quality as no significant relationship was found
 344 between H/D ratio and nurse-plant size or distance to the nurse-plant edge (Fig. S2).

345 *Relationship between size and age of nurse plant*

346 Finally, we found a significant second polynomial relationship between nurse-plant age and size
 347 (Predicted nurse-plant size= $0.13+0.55*Age-0.017*Age^2$; Fig. S3). At approximately 16 years old,
 348 nurse-plant size peaked and then started to decrease (Fig. S3).

349 **DISCUSSION**

350 Our results show that both components of plant facilitation (quantity and quality of oak recruits)
351 improved with increasing nurse-plant size. However, contrary to other studies that attributed the
352 increased facilitative effect mostly to the amelioration of abiotic conditions under larger nurse
353 plants (Pugnaire et al., 1996b; Tewksbury & Lloyd, 2001), we pose that, in ungulate-dominated
354 environments, herbivore stress strongly mediates in the observed effect.

355 For both levels of herbivore stress, we found that, quantitatively, the recruit density increased with
356 nurse-plant size. Only large nurse plants allowed plant facilitation, with greater regeneration
357 densities underneath as compared to microsites located outside the nurse plant. This agrees with
358 previous studies showing that plant facilitation mostly occur under nurse plants of sufficient size
359 (Pugnaire et al., 1996b; Drezner, 2006; Smit et al., 2006; Perea et al., 2016). However, our results
360 highlight that the net effect of nurse-plant size is strongly mediated by the level of the biotic stress.
361 Thus, at high levels of herbivore stress, the probability of facilitation occurrence (i.e., higher recruit
362 density underneath than outside the nurse plant) started in nurse plants of approximately half size
363 (2.4 m crown diameter) than in those located at low levels of stress (starting at 4.7 m crown
364 diameter). These results reveal that, even though nurse-plant size is important in plant facilitation
365 at low and high biotic stress, the magnitude of the effect was strongly mediated by the intensity of
366 herbivory. Therefore, facilitation was more pronounced and required smaller nurse-plant sizes at
367 higher levels of biotic stress. These findings support the predictions of the SGH as the net effect
368 of plant facilitation increased for greater levels of biotic stress but facilitative effects were strongly
369 dependent on nurse-plant traits. This is also consistent with the growing evidence indicating that
370 plant facilitation is strongly dependent on the physical and chemical traits of the nurse plant
371 (Bruno, Stachowicz, & Bertness, 2003; Baraza et al., 2006; Perea & Gil, 2014b). Accordingly, we

372 found a trait-based threshold for nurse plants to provide facilitation but, importantly, that threshold
373 was strongly dependent on the stress level. This trait-based framework may explain why some
374 studies found no facilitation at extreme biotic conditions (Smit et al., 2007; Vandenberghe et al.,
375 2009) where plant traits are disrupted or thresholds are not reached. Our results suggest that
376 facilitation is highly context-dependent with a strong hierarchical effect of plant traits over the
377 biotic stress level. These ideas are indeed consistent with Maestre et al. (2009), who have sought
378 to place the complexity and species-specificity of plant-plant interactions along biotic and abiotic
379 stress gradients.

380 Similarly, recruit quality improved with increasing nurse-plant size. Thus, larger nurse plants
381 sheltered recruits with proportionally higher H/D ratios. This ratio, a proxy for plant growth quality
382 used in herbivore-dominated environments (Martínez & López-Portillo, 2003), showed a strong
383 correlation with the measured browsing damage index (Perea et al., 2015). Therefore, as
384 hypothesized, larger nurse plants protected recruits better from herbivory (Hay, 1986; Gómez et
385 al., 2001; Perea et al., 2016) and hence allowed beneficiaries to grow well, with balanced H/D
386 ratios.

387 In addition, this pattern of greater qualitative facilitation under larger nurse plants was confirmed
388 by the strong positive relationship between recruit quality and distance to nurse-plant edge.
389 Furthermore, quantitative results on the distribution of plants under the nurse plants reflected that
390 recruits were distributed differently depending on the herbivore stress. Thus, at higher herbivore
391 stress, recruits were found, on average, at greater distances from the nurse-plant edge. Other studies
392 have shown a similar pattern of recruit clustering around the nurse-plant base (Valiente-Banuet et
393 al., 1991; Drezner, 2006), which has been mostly attributed to the fact that abiotic stress decreases
394 towards the nurse-plant base (Castellanos et al., 1999). In environments with high herbivore stress,

395 this clustered distribution around the plant base could also be attributed to the higher probability
396 of recruits being browsed at shorter distances from the nurse-plant edge. Thus, both the qualitative
397 and the quantitative facilitative effect notably increased with increasing distance to the nurse-plant
398 edge (i.e., with larger nurse plants), which indicates the importance of considering this parameter
399 in the facilitation process in herbivore-dominated environments.

400 Interestingly, the relative importance of the qualitative component varied with recruit ontogenetic
401 stage. Thus, for saplings, the qualitative component was more pronounced for increasing distances
402 to nurse-plant edge whereas for seedlings the qualitative component was not associated with
403 distance to nurse-plant edge. This is, indeed, consistent with previous studies that detected a lower
404 browsing damage on seedlings compared to saplings in ungulate-dominated environments (Rao et
405 al. 2003; López-Sánchez et al., 2014; Perea et al., 2016). This lack of relationship between
406 seedlings quality and distance to edge is probably related to the difficulty to detect browsing
407 damage on them, as often herbivory results in the elimination of the aerial part of the recruit.
408 Therefore, for seedlings, the biotic stress can be evaluated through the quantitative component
409 while it is difficult to assess the effect on quality (we found no significant relationship between
410 herbivory damage and H/D for this plant category). He and Bertness (2014) have also argued that
411 exceptions to the SGH may occur when the life-history stage of beneficiary species shifts across
412 different stress conditions. For instance, several studies highlighted that qualitative facilitation
413 (e.g. growth) may be most important at intermediate stress levels (Smit et al., 2007; Holmgren &
414 Scheffer, 2010), whereas others show that quantitative facilitation (e.g. density of regeneration) is
415 greater at highest stress levels (Perea et al., 2017). However, here, we showed that the effect of the
416 same stress conditions may be better evaluated by either the qualitative and quantitative
417 components of plant facilitation at different ontogenetic stages. These results reveal the importance

418 of considering both components (quantity and quality) in plant facilitation studies to fully
419 understand how species interactions change at increasing levels of stress.

420 *Size-age relationship in plant facilitation*

421 As expected, plant age and size were positively correlated until plants reached 16 years old when
422 plants begin to decrease in size. This is consistent with observations of Hobbs and Mooney (1986)
423 who described that after maturity is reached, at approximately 9 years old, a decline in biomass
424 occurs until canopy collapses and shrub dies. Zavaleta and Kettlely (2006) already pointed out that
425 *Baccharis* are short-living shrubs that contribute to plant succession in oak-dominated systems.
426 Following the age-size relationship described here (Fig. S1), facilitation started to occur at the
427 approximate age of 7 years (equivalent to ~2.5 m of crown diameter at high herbivory level).
428 However, it remains elusive whether plant facilitation is determined by nurse-plant age, size or a
429 combination of both. It seems reasonable that plant size plays a major role for protection against
430 herbivores rather than age whereas, for abiotic stress, amelioration with age may play a more
431 significant role due to an increase of soil nutrient availability for older plants (Pugnaire et al.,
432 1996b). Here, we found a different pattern of beneficiary plant distribution depending on the
433 herbivore stress. For low levels of herbivory, plants were found on average closer to the nurse-
434 plant edge whereas for high levels of herbivory plants were found further inside. This confirms,
435 that plant size rather than age plays a major role in protecting beneficiaries against herbivory and
436 it also contributes to explain why size threshold for plant facilitation varied with increasing biotic
437 stress level. However, further studies should specifically address size and age effect along biotic
438 and abiotic stress gradients by controlling one of the variables and experimentally manipulating
439 the other. Further work to assess the extent to which the complex interplay of relationships
440 uncovered here occurs in other systems of concurrent biotic and abiotic stress is warranted.

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449 **REFERENCES**

450 Armas, C., & Pugnaire, F. I. (2009). Ontogenetic shifts in interactions of two dominant shrub
451 species in a semi-arid coastal sand dune system. *Journal of Vegetation Science*, 20, 535-546.

452 Baraza, E., Zamora, R., & Hódar, J. A. (2006). Conditional outcomes in plant-herbivore
453 interactions: neighbours matters. *Oikos*, 113, 148–156.

454 Bartolome, J. W., McClaran, J. P., Allen-Diaz, B. H., Dunne, J., Ford, L.D., Standiford, R.B.,
455 McDougald, N.K., & Forero, L.C. (2002). Effects of fire and browsing on regeneration of blue
456 oak. Pages 281-286 in R.B. Standiford, D. McCreary, and K.L. Purcell, editors. Proceedings of the
457 fifth symposium on oak woodlands: Oaks in California's changing landscape; USDA Forest
458 Service General Technical Report PSW-GTR-184.

459 Bertness, M. D., & Callaway, R. (1994). Positive interactions in communities. *Trends in Ecology*
460 *and Evolution*, 9, 191–193.

461 Bertness, M. D., & Leonard, G. (1997). The role of positive interactions in communities: Lessons
462 from the intertidal. *Ecology*, 78, 1978-1989.

463 Brooker, R. W., Maestre, F. T., Callaway, R. M., Lortie, C. L., Cavieres, L. A., Kunstler, G., ... &
464 Armas, C. (2008). Facilitation in plant communities: the past, the present, and the future. *Journal*
465 *of Ecology*, 96, 18–34.

466 Bruno, J. F., Stachowicz, J. J., & Bertness, M. D. (2003). Inclusion of facilitation into ecological
467 theory. *Trends in Ecology and Evolution*, 18, 119–125.

468 Burnham, K. P., & Anderson, D. R. (2002). *Model selection and multimodel inference: a practical*
469 *information-theoretic approach*. New York: Springer-Verlag.

470 Callaway, R. M. (1992). Effect of shrubs on recruitment of *Quercus douglasii* and *Quercus lobata*
471 in California. *Ecology*, 73, 2118–2128.

472 Callaway, R. M., & Davis, F. W. (1998). Recruitment of *Quercus agrifolia* in central California:
473 the importance of shrub-dominated patches. *Journal of Vegetation Science*, 9, 647–656.

474 Callaway, R. M., Kikodze, D., Chiboshvili, M., & Khetsuriani, L. (2005). Unpalatable plants
475 protect neighbors from grazing and increase plant community diversity. *Ecology*, 86, 1856–1862.

476 Callaway, R. M. (2007). Positive interactions and interdependence in plant communities. Springer,
477 Dordrecht, The Netherlands. ISBN 978-1-4020-6223-0 (HB).

478 Carmen, W. J. (2004). Behavioral ecology of the California scrub jay (*Aphelocoma californica*): a
479 non-cooperative breeder with close cooperative relatives. *Studies in Avian Biology*, 28:1–100.

480 Castellanos, A. E., Tinoco-Ojanguren, C., & Molina-Freaner, F. (1999). Microenvironmental
481 heterogeneity and space utilization by desert vines within their host trees. *Annals of Botany*, 84,
482 145–153.

483 Côté, S. D., Rooney, T. P., Tremblay, J. P., Dussault, C., & Waller, D. (2004). Ecological impacts
484 of deer overabundance. *Annual Review of Ecology, Evolution, and Systematics*, 35, 113–147.

485 Domínguez, M. T., Pérez-Ramos, I. M., Murillo, J. M., & Marañón, T. (2015). Facilitating the
486 afforestation of Mediterranean polluted soils by nurse shrubs. *Journal of environmental*
487 *management*, 161, 276-286.

488 Drezner, T. D. (2006). Plant facilitation in extreme environments: the non-random distribution of
489 saguaro cacti (*Carnegiea gigantea*) under their nurse associates and the relationship to nurse
490 architecture. *Journal of Arid Environments*, 65, 46–61.

491 Franco, A. C., & Nobel, P. S. (1988). Interactions between seedlings of *Agave deserti* the nurse
492 plant *Hilaria Rigida*. *Ecology*, 69, 1731–1740.

493 Gómez, J. M., Hódar, J. A., Zamora, R., Castro, J., & García, D. (2001). Ungulate damage on Scots
494 pines in Mediterranean environments: effects of association with shrubs. *Canadian Journal of*
495 *Botany*, 79, 739–746.

496 Gómez-Aparicio, L., Zamora, R., Castro, J., & Hódar, J. A. (2008). Facilitation of tree saplings by
497 nurse plants: Microhabitat amelioration or protection against herbivores? *Journal of Vegetation*
498 *Science*, 19(2), 161–172.

499 Gómez-Aparicio, L., Zavala, M. A., Bonet, F. J., & Regino Zamora, R. (2009). Are pine
500 plantations valid tools for restoring Mediterranean forests? An assessment along gradients of
501 climatic conditions, stand density and distance to seed sources. *Ecological Applications*, 19, 2124–
502 2141.

503 Gordon, I. J., Hester, A. J., & Festa-Bianchet, M. (2004). The management of wild large herbivores
504 to meet economic, conservation and environmental objectives. *Journal of Applied Ecology*, 41,
505 1021–1031.

506 Griffin, J. R. (1977). Oak woodland. In: Barbour, Michael G.; Malor, Jack, eds. *Terrestrial*
507 *vegetation of California*. New York: John Wiley and Sons: 383–415.

508 Hay, M. E. (1986). Associational Plant Defenses and the Maintenance of Species Diversity:
509 Turning Competitors Into Accomplices. *The American Naturalist*, 128, 617–641

510 He, Q., & Bertness, M. D. (2014). Extreme stresses, niches, and positive species interactions along
511 stress gradients. *Ecology*, 95, 1437–1443.

512 Hobbs, R. J., & Mooney, H. A. (1986). Community changes following shrub invasion of
513 grassland. *Oecologia*, 70, 508–513.

514 Holmgren, M., & Scheffer, M. (2010). Strong facilitation in mild environments: the stress gradient
515 hypothesis revisited. *Journal of Ecology*, 98, 1269–1275.

516 Kamimura, K., & Shiraishi, N. (2007). A review of strategies for wind damage assessment in
517 Japanese forests. *Journal of Forest Research*, 12(3), 162–176.

518 Lambert, D. (1992). Zero-inflated Poisson regression, with an application to defects in
519 manufacturing. *Technometrics*, 34, 1–14.

520 Leiva, M. J., Mancilla-Leyton, J. M., & MartínVicente, Á. (2015). Differences in the facilitative
521 ability of two Mediterranean shrubs on holm-oak seedling recruitment in Mediterranean savanna-
522 forest ecosystems. *Ecological Engineering*, 82, 349-354.

523 López-Sánchez, A., Schroeder, J., Roig, S., Sobral, M., & Dirzo, R. (2014). Effects of cattle
524 management on oak regeneration in northern Californian Mediterranean oak woodlands. *PloS*
525 *one*, 9(8), e105472.

526 López-Sánchez, A., Perea, R., Dirzo, R., & Roig, S. (2016). Livestock vs. wild ungulate
527 management in the conservation of Mediterranean dehesas: Implications for oak regeneration.
528 *Forest Ecology and Management*, 362, 99–106

529 Maestre, F. T., Valladares, F., & Reynolds, J. F. (2005). Is the change of plant-plant interactions
530 with abiotic stress predictable? A meta-analysis of field results in arid environments. *Journal of*
531 *Ecology*, 93, 748–757.

532 Maestre, F. T., Callaway, R. M., Valladares, F., & Lortie, C. J. (2009) Refining the stress-gradient
533 hypothesis for competition and facilitation in plant communities. *Journal of Ecology*, 97, 199–205

534 Martínez, A. J., & López-Portillo, J. (2003). Growth and architecture of small Honey Mesquites
535 under jackrabbit browsing: Overcoming the disadvantage of being eaten. *Annals of Botany*, 92,
536 365–375,

537 McBride, J. R. (1974). Plant succession in the Berkeley Hills, California. *Madroño*, 22, 317–380.

538 McBride, J. R., & Heady, H. F. (1968). Invasion of grassland by *Baccharis pilularis* DC. *Journal*
539 *of Range Management*, 21, 106–108

540 Palmer, B. (2003). Committee for Dish recreation use issues final report. *Stanford News Service*.
541 Stanford, CA.

542 Perea, R., & Gil, L. (2014a). Shrubs facilitating seedling performance in ungulate-dominated
543 systems: biotic versus abiotic mechanisms of plant facilitation. *European Journal of Forest*
544 *Research*, 133, 525–534

545 Perea, R., & Gil, L. (2014b). Tree regeneration under high levels of wild ungulates: The use of
546 chemically vs. physically-defended shrubs. *Forest Ecology and Management*, 312, 47–54

547 Perea, R., Perea García-Calvo, R., Díaz-Ambrona, C. G., & San Miguel, A. (2015). The
548 reintroduction of a flagship ungulate *Capra pyrenaica*: Assessing sustainability by surveying
549 woody vegetation. *Biological Conservation*, 181, 9–17.

550 Perea, R., López-Sánchez, A., & Roig, S. (2016). The use of shrub cover to preserve Mediterranean
551 oak dehesas: a comparison between sheep, cattle and wild ungulate management. *Applied*
552 *Vegetation Science*, 19, 244–256.

553 Perea, R., López-Sánchez, A., & Dirzo, R. (2017). Differential tree recruitment in Californian oak
554 savannas: Are evergreen oaks replacing deciduous oaks? *Forest Ecology and Management*, 399,
555 1–8.

556 Perea, R., Fernandes, G. W., & Dirzo, R. (2018). Embryo size as a tolerance trait against seed
557 predation: Contribution of embryo-damaged seeds to plant regeneration. *Perspectives in Plant*
558 *Ecology, Evolution and Systematics* 31: 7-16.

559 Pesendorfer, M. B., & Koenig, W. D. (2016). The effect of within-year variation in acorn crop size
560 on seed harvesting by avian hoarders. *Oecologia*, 181(1), 97-106.

561 Pugnaire, F. I., Haase, P., & Puig de Fabregas, J. (1996a). Facilitation between higher plant species
562 in a semiarid environment. *Ecology*, 77(5), 1420–1426

563 Pugnaire, F. I., Haase, P., Puig de Fábregas, J., Cueto, M., Incoll, L. D., & Clack, S. C. (1996b).
564 Facilitation and succession under the canopy of the leguminous shrub, *Retama sphaerocarpa*, in a
565 semi-arid environment in south-east Spain. *Oikos*, 76, 455–464.

566 R Core Team (2018). R: A language and environment for statistical computing. R Foundation for
567 Statistical Computing, Vienna, Austria. URL <https://www.R-project.org/>. Accessed 23 April 2018

568 Rao, S. J., Iason, G. R., Hulbert, I. A. R., Elston, D. A., & Racey, P. A. (2003). The effect of
569 sapling density, heather height and season on browsing by mountain hares on birch. *Journal of*
570 *Applied Ecology*, 40, 626–638.

571 Sampson, A. W., & Jespersen, B. S. (1963). California range brushlands and browse plants.
572 Berkeley, CA: University of California, Division of Agricultural Sciences, California Agricultural
573 Experiment Station, Extension Service.

574 Schelhaas, M. J. (2008). The wind stability of different silvicultural systems for Douglas-fir in the
575 Netherlands: a model-based approach. *Forestry*, 81(3), 399–414.

576 Schupp, E. W., Jordano, P., & Gómez, J. M. (2010). Seed dispersal effectiveness revisited: a
577 conceptual review. *New Phytologist*, 188, 333–353.

578 Schwinning, S., & Weiner, J. (1998). Mechanisms determining the degree of size-asymmetry in
579 plant competition. *Oecologia*, 113, 447–455.

580 Smit, C., den Ouden, J., & Muller-Scharer, H. (2006). Unpalatable plants facilitate tree sapling
581 survival in wooded pastures. *Journal of Applied ecology*, 43, 305–312.

582 Smit, C., Vandenberghe, C., Ouden den, J., & Müller-Schärer, H. (2007). Nurse plants, tree
583 saplings and grazing pressure: changing facilitation along a biotic environmental gradient.
584 *Oecologia*, 152, 265–273.

585 Smit, C., den Ouden, J., & Díaz, M. (2008). Facilitation of holm oak recruitment by shrubs in
586 Mediterranean open woodlands. *Journal of Vegetation Science*, 19, 193–200.

587 Smit, C., Rietkerk, M., & Wassen, M. (2009). Inclusion of biotic stress (consumer pressure) alters
588 predictions from the stress gradient hypothesis. *Journal of Ecology*, 97, 1215–1219.

589 Smither-Kopperl, M. (2016). Plant Guide for coyotebrush (*Baccharis pilularis*). USDA-Natural
590 Resources Conservation. Service, Lockeford Plant Materials Center, Lockeford, CA.

591 Soliveres, S., Eldridge, D. J., Maestre, F. T., Bowker, M. A., Tighe, M., & Escudero, A. (2011).
592 Microhabitat amelioration and reduced competition among understory plants as drivers of
593 facilitation across environmental gradients: Towards a unifying framework. *Perspectives in Plant*
594 *Ecology, Evolution and Systematics*, 13, 247–258.

595 Steinberg, P. D. (2002). *Baccharis pilularis*. In: Fire Effects Information System, [Online]. U.S.
596 Department of Agriculture, Forest Service, Rocky Mountain Research Station, Fire Sciences
597 Laboratory (Producer).

598 Tewksbury, J. J., & Lloyd, J. D. (2001). Positive interactions under nurse-plants: spatial scale,
599 stress gradients and benefactor size. *Oecologia*, 127, 425–434.

600 Tirado, R., & Pugnaire, F. I. (2005). Community structure and positive interactions in constraining
601 environments. *Oikos*, 111, 437–444.

602 Valiente-Banuet, A., Bolongaro-Crevenna, A., Briones, O., Ezcurra, E., Rosas, M., Nuñez, H., ...
603 & Vazquez, E. (1991). Spatial relationships between cacti and nurse shrubs in a semi-arid
604 environment in central Mexico. *Journal of Vegetation Science*, 2, 15–20.

605 Valiente-Banuet, A., & Ezcurra, E. (1991). Shade as a cause of the association between the cactus
606 *Neobuxbaumia tetetzo* and the nurse plant *Mimosa luisana* in the Tehuacan Valley, Mexico. *The*
607 *Journal of Ecology*, 961–971.

608 Valladares, F., & Pearcy, R. W. (2002). Drought can be more critical in the shade than in the sun:
609 a field study of carbon gain and photoinhibition in a Californian shrub during a dry El Niño year.
610 *Plant Cell and Environment*, 25, 749–759

611 Vandenberghe, C., Smit, C., Pohla, M., Buttlera, A., & Freléchoux, F. (2009) Does the strength
612 of facilitation by nurse shrubs depend on grazing resistance of tree saplings? *Basic and Applied*
613 *Ecology*, 10, 427–436.

614 Vospernik, S., Monserud, R. A., & Sterba, H. (2010). Do individual-tree growth models correctly
615 represent height: diameter ratios of Norway spruce and Scots pine? *Forest ecology and*
616 *management*, 260, 1735–1753.

617 Vuong, Q. H. (1989). Likelihood ratio tests for model selection and non-nested hypotheses,
618 *Econometrica*, 57, 307–33.

619 Zavaleta, E. S., & Kettley, L. S. (2006). Ecosystem change along a woody invasion
620 chronosequence in a California grassland. *Journal of Arid Environments*, 66, 290–306.

622 **SUPPLEMENTARY MATERIAL**

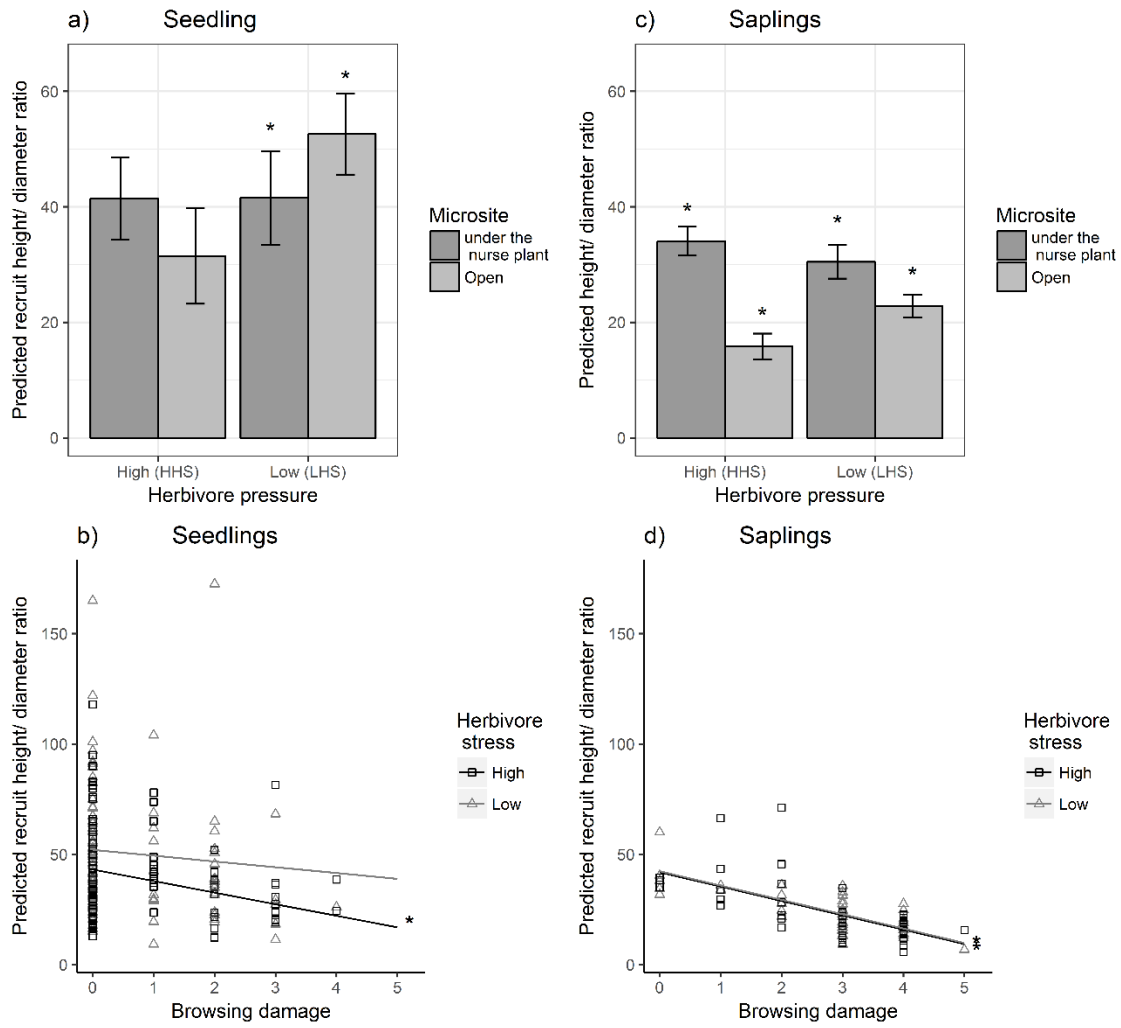
623 **Figure legends**

624 **Fig. S1.** Histograms show microsite differences in predicted mean H/D ratio for seedlings (a) and
625 saplings (c) at high (HHS) and low (LHS) herbivore stress. Linear regression represent the
626 relationship of H/D ratio with browsing damage for seedlings (b) and saplings (d) at HHS and
627 LHS herbivore stress. The significance is reported at each herbivore stress site (* $p < 0.05$).

628 **Fig. S2. a)** Predicted H/D ratio for seedlings located under the nurse plant microsite in relation to
629 nurse-plant size at high (HHS; black lines) and low (LHS; grey lines) herbivore stress. b) Predicted
630 seedling H/D ratio in relation to the distance to the nurse-plant edge for both microsities (open and
631 under nurse plant) and for high (HHS) and low (LHS) herbivore stress. Square and triangles
632 represent seedlings in HHS and LHS, respectively. Color scale represents browsing damage
633 recorded for each individual sapling (ns = not significant).

634 **Fig S3.** Predicted relationship between nurse-plant size and its minimum estimated age (black
635 line). Black squares represent each of the 108 nurse plants measured in the high herbivore stress
636 site.

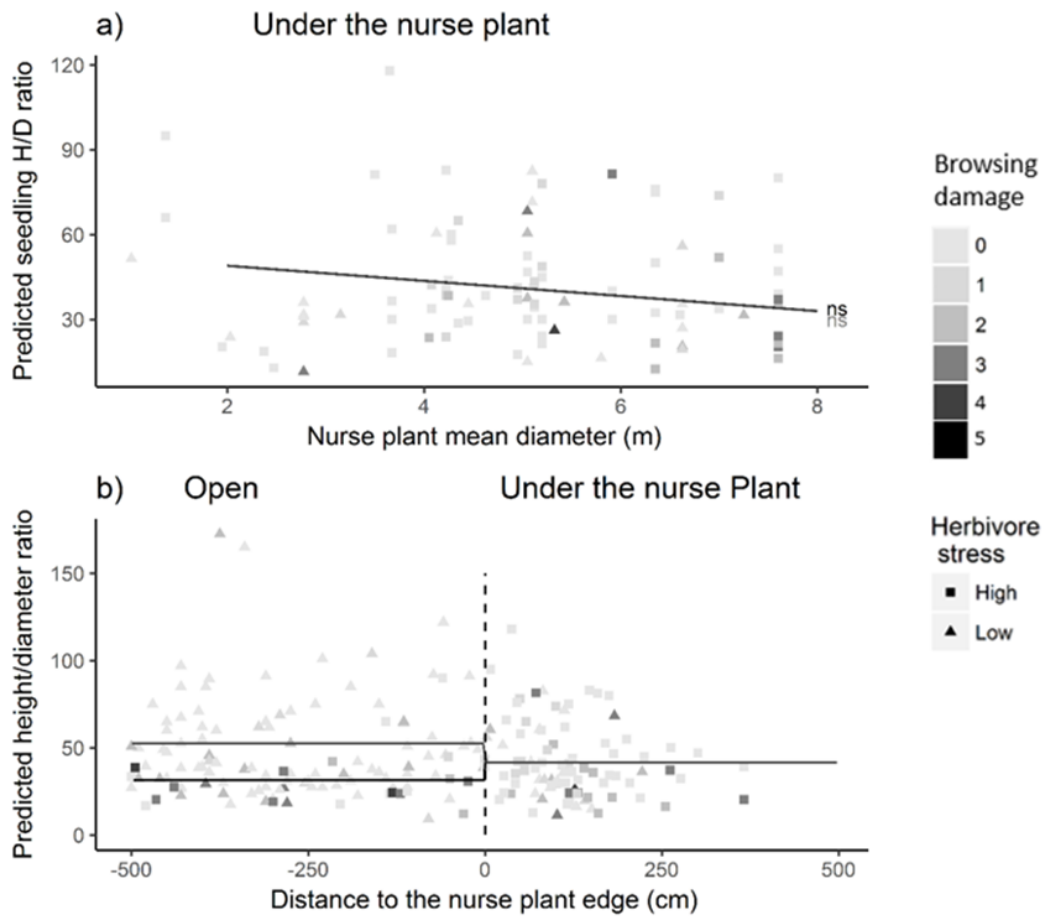
637 **Figure S1**



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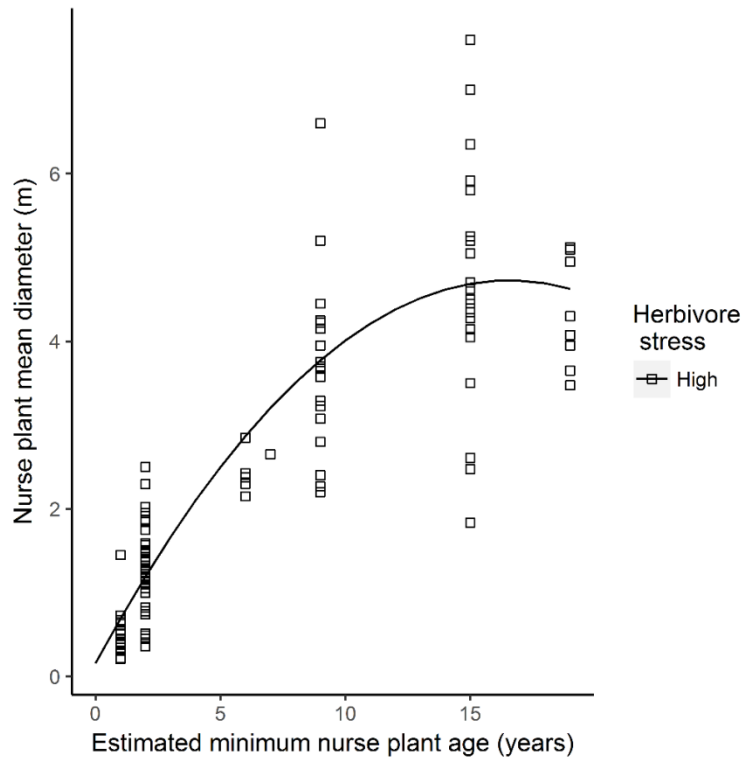
640 **Figure S2**



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643 **Figure S3**



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